Tampa Bay Benthic Monitoring Program: Status of Hillsborough Bay: 1993-1998

Stephen A. Grabe Environmental Supervisor

David J. Karlen Environmental Scientist II

Christina M. Holden Environmental Scientist I

Barbara Goetting Environmental Specialist I

Thomas Dix Environmental Scientist II

Environmental Protection Commission of Hillsborough County

November 2002

Environmental Protection Commission of Hillsborough County

Dr. Richard Garrity

Executive Director

Tom Cardinale

Acting Director

Environmental Resources Management Division

ACKNOWLEDGEMENTS

Funding was provided by the Tampa Bay Estuary Program (1993-1998), the Environmental Protection Commission of Hillsborough County, and the Phosphate Severance Tax. The USEPA/Gulf Breeze provided additional laboratory support for the 1993 and 1997 surveys. Tom Ash, Glenn Lockwood, Richard Boler, and Eric Lesnett assisted with field collections and instrument calibration. Sediment chemical analyses were performed by Joseph Barron and Steven Perez. Sediment particle size analysis was provided by Manatee County's Environmental Management Department. Laboratory assistance was provided by a plethora of temporary employees over the years. D. Camp (Crustacea), R. Heard (Peracarida), S. LeCroy (Amphipoda), W. Lyons (Mollusca), M. Milligan (Oligochaeta), T. Perkins (Polychaeta), W. Price (Mysidacea), K. Strasser (Paguroidea), J.S. Harrison (Pinnotheridae), and H.K. Dean (Sipuncula) verified/identified specimens for us. Holly Greening (Tampa Bay Estuary Program) reviewed an earlier draft of this document.

EXECUTIVE SUMMARY

The Environmental Protection Commission of Hillsborough County (EPCHC) has been collecting sediment samples on an annual (summer) basis in Hillsborough Bay since 1993 as part of a bay-wide monitoring program developed by the Tampa Bay National Estuary Program. These samples are analyzed for the composition and abundance of the animals living in and on the sediments ("benthos") as well as for chemical contaminants (metals, pesticides *etc.*). The original objectives of this program were to discern the "health"—or "status"-- of the bay's sediments based upon both chemistry and biology.

The Tampa Bay Estuary Program (formerly the Tampa Bay National Estuary Program) and the USEPA have provided partial support for this monitoring program.

This report summarizes data collected during 1993-1998 from the Hillsborough Bay segment of Tampa Bay.

1- 139 locations (19 to 29 per year) were sampled during late summer/early fall "Index Period" from 1993 to 1998.

2- Near-bottom water temperatures during 1994 were similar to 1997; 1998 and 1996 water temperatures were similar to 1997.

3- Near bottom salinities were generally highest in 1997 and lowest during the 1995 and 1998 sampling periods.

4- Salinities were generally within the polyhaline (18-30 ppt) zone. Mesohaline (5-18 ppt) conditions were observed along the periphery of the bay during 1995 and, to a greater extent, during 1998.

5-Near-bottom dissolved oxygen concentrations were less than 2 parts per million (hypoxia) in more than 30% of samples collected during 1998 and in less than 10% of 1996 samples.

4

6- Hypoxia was most often observed immediately downstream of the Hillsborough River, between the Interbay peninsula and the Davis Islands.

7- Sediments in Hillsborough Bay were predominantly sandy, although muds are widely distributed in western Hillsborough Bay.

8- At the deeper sites in Hillsborough Bay, the percentage of fine-grained (mud and silt) sediments was higher and dissolved oxygen concentrations lower.

9- A composite index of the chemical contamination of Hillsborough Bay sediments suggested that, in any year surveyed to date, less than 5% of the of the samples had a high likelihood of being toxic to aquatic life. Chemically degraded sediments were found in the upper portions of Hillsborough Bay, including McKay Bay, East Bay, and the Lower Hillsborough River.

10- The percentage of samples "degraded" by metals and hydrocarbons (oils, automotive combustion products, etc.) ranged from 0 to 4.8; the percentage degraded by PCBs ranged from 0 to 3.7% in any year. Additionally, as many as 10% of the samples in any year were likely to be toxic for chlordane.

11- Numerically abundant benthic species included the clam *Mysella planulata,* several segmented worms (*Monticellina dorsobranchialis*, *Carazziella hobsonae, Prionospio perkinsi*), non-segmented worms, and the amphipod crustacean ("scud") *Ampelisca holmesi.*

12-Numerical dominants differed both by year and salinity zone.

13- The variety of animals in any sample varied over years. In 1995, the median numbers of taxa per sample was less than 20 whereas in 1994 it was more than 30. During 1995, almost 20% of the samples contained no living animals.

14- Tampa Bay Benthic Index (a composite measure of the "health" of the communities of bottom dwelling organisms) scores were generally higher during 1993, 1995 and 1996 than during 1997 and 1998. Degraded habitat was found at approximately 30% of the sites sampled during 1995—*vs*. less than 15% of sites in other years. Degraded benthic habitat was generally located immediately downstream of the Lower Hillsborough River, between the Interbay Peninsula and Davis Island, in the northeastern, industrial area of Hillsborough Bay.

15- Benthic Index scores were positively associated with dissolved oxygen and negatively associated with increasing percentages of fine-grained sediments (muds, silts), sediment contaminant level, and depth. The Benthic Index was not affected by salinity.

16- The overall structure of the community of sediment dwelling animals (benthos) was not clearly explained by patterns or trends in physical factors generally shown to be linked to benthic community structure (*e.g*., salinity and sediment type). The linkage between the structure of the benthic community and structure based upon physical (*e.g.,* sample depth, sediment type) and chemical contaminants (*e.g*., metals, pesticides) was generally weak except when yearly survey averages were considered

17- The benthic community experienced more pronounced shifts in structure from 1996 to 1997 and from 1997 to 1998 than during other years. Three metals (arsenic, cadmium, and chromium) each showed a marked increase in mean concentration from 1996 to 1997. From 1997 to 1998 arsenic concentrations declined approximately 75% while chromium increased almost 5 fold. Benthic organisms which contributed primarily to this shift in community structure from 1996 to 1997 included the clam *Mysella planulata*, the segmented worm *M. dorsobranchialis*, and the "scud" *Ampelisca holmesi*. Shifts in structure from 1997 to 1998 were primarily affected by alterations in the abundance of the segmented worms *M. dorsobranchialis, Prionospio perkinsi*, and the scud *A. holmesi*.

18- Hillsborough Bay appears to be affected more by subnominal dissolved oxygen concentrations than the Louisianian Province (northern Gulf of Mexico south to Tampa Bay) as a whole. Degradation of benthic habitat, however, appears to be markedly higher in the Louisianian Province than in Hillsborough Bay.

19- The composition of the benthos appears to have undergone changes since the 1960s and 1970s. Changes were observed in the most frequently occurring species within three taxonomic groups: segmented worms, clams and snails, and one group of amphipods ("scuds"). Although interannual variations in population size and differences in sampling locations could explain some of these differences, such differences may also reflect changes in habitat quality over the past 30 years.

TABLE OF CONTENTS

LIST OF TABLES

LIST OF FIGURES

LIST OF FIGURES (continued)

conditions, and \geq 4 ppm indicates "healthy" conditions.

LIST OF FIGURES (continued)

Figure 26. Association between the Tampa Bay Benthic Index and %SC in Hillsborough Bay, 1993-1998. TBBI scores <4.6 indicates "degraded" habitat and scores >4.6 indicate "healthy" habitat. Sand and mud fractions are demarcated by 25.95%SC 26

Figure 27. Association between the Tampa Bay Benthic Index and the apparent redox potential discontinuity layer (RPD) in Hillsborough Bay, 1993-1998. TBBI scores <4.6 indicate "degraded" habitat and scores >4.6 indicate "healthy" habitat. 27

Figure 28. Association between the Tampa Bay Benthic Index and the composite PEL quotient in Hillsborough Bay, 1993-1998. TBBI scores <4.6 and and scores >4.6 indicate "healthy" habitat. PEL quotients >0.34 indicate "degraded" sediments and PEL quotients <0.05 indicate "clean" sediments**.** 27

Figure 29. Association between the Tampa Bay Benthic Index and sample depth in Hillsborough Bay, 1993-1998. TBBI scores <4.6 indicate "degraded" habitat and scores >4.6 indicate "healthy" habitat. 28

.

.

Figure 30. Association between the Tampa Bay Benthic Index and salinity in Hillsborough Bay, 1993-1998. TBBI scores <4.6 indicate "degraded" habitat and scores >4.6 indicate "healthy" habitat. Salinity zones are demarcated as: oligohaline $(<$ 5 ppt), mesohaline (5-18 ppt), and polyhaline (18-30 ppt). 27

Figure 31. Dendrogram depicting the similarity of sites in Hillsborough Bay 1993- 1998 ($4th$ root transformed abundance; Bray-Curtis similarity; group-average clustering). 31

Figure 32. Dendrogram depicting the similarity of the 50 most abundant taxa in Hillsborough Bay 1993-1998 (standardized abundance; Bray-Curtis similarity; group-average clustering).. 32

Figure 33. MDS representation of benthic community structure in Hillsborough Bay, 1993-1998, by year. 38

LIST OF FIGURES (continued)

Figure 34. "Bubble" plots of chlordane, PCBs, Al, and Cu superimposed over the MDS plot depicting benthic community structure, by year, in Hillsborough Bay 1995-1998. 39

Figure 35. MDS plot of "average" benthic community structure by year and mean concentrations of As, Cd, and Cr by year, Hillsborough Bay, 1993-1998. Lines delineate temporal trend. 39

Figure 36. Mean annual densities (# m-2) of *Mysella planulata, Ampelisca holmesi*, Tubificidae, *Monticellina dorsobranchialis*, and *Prionospio perkinsi* in Hillsborough Bay 1993-1998 superimposed over the MDS structure (*cf*. Figure 33). 41

LIST OF APPENDICES

INTRODUCTION

The Environmental Protection Commission of Hillsborough County (EPCHC) has been collecting benthos and sediment samples on an annual basis (late summer to early fall) in Hillsborough Bay since 1993 as part of a bay-wide monitoring program developed by theTampa Bay National Estuary Program (1996). The original objectives of this program were to discern the "health"—or "status"-- of the bay's sediments by developing a Benthic Index for Tampa Bay and by evaluating sediment quality by applying Sediment Quality Assessment Guidelines (SQAGs). The Tampa Bay Estuary Program (formerly the Tampa Bay National Estuary Program) and the USEPA have provided partial support for this monitoring program.

This report summarizes data collected during 1993-1998 from the Hillsborough Bay segment of Tampa Bay.

METHODS

Field Collection and Laboratory Procedures: A total of 139 stations (19 to 29 per year) were sampled during a late summer through early fall "Index Period" from 1993 through 1998 (Appendix A). Sample locations were randomly selected from computer- generated coordinates. Benthic samples were collected using a Young grab sampler (Figure 1) following the field protocols outlined in Courtney *et al*. (1993). Laboratory procedures followed the protocols set forth in Courtney *et al*. (1995).

Figure 1. Young grab sampler used to collect sediment and benthic samples.

Data Analysis: Species richness, Shannon-Wiener diversity, and Evenness were calculated using PISCES Conservation Ltd.'s (2001) "Species Diversity and Richness II" software. Descriptive statistics, the Tampa Bay Benthic Index (TBBI), regression analysis, and the Kolmogorov-Smirnov "two-sample" test (used to compare frequency distributions by year), and graphs were generated using SYSTAT 10 (SSPS Inc. 2000). Sediment status was assessed by comparing measured concentrations with the Predicted Effects Level (PEL) developed for Florida sediments by McDonald Environmental Sciences Ltd. (1994). A composite PEL quotient (based upon PAHs, PCBs and metals) >0.34 and TBBI scores <4.6 were considered to be "degraded"—*i.e*., having a high likelihood of being associated with toxic sediments (MacDonald *et al*. 2002). Maps were generated using GIS Arcview ver. 3.2 (ESRI 1999).

Principal Components Analysis (PCA) (PRIMER-E Ltd. 2001) was used to examine the resemblance of the Hillsborough Bay sites, by year. Hydrographic (temperature, salinity, dissolved oxygen) and sediment (percent silt+clay [%SC]) variables were normalized prior to anlaysis. The objective of this ordination is to reduce the multiple variables into a lower dimensional (2) "map" based upon the percentage of the total variance explained (principal component) (Clarke & Warwick 2001). "Bubble" plots were superimposed over the ordination diagram representing the variables with the highest "loading" (*i.e*., the "importance" of a particular variable to that principal component [PC]; Johnson & Wichern 1988) in the first two PCs to facilitate interpretation of the ordination.

Non-metric Multidimensional Scaling (MDS) is another ordination technique in which rank similarities of a large number of variables are expressed as a two-dimensional map (Clarke & Warwick 2001). In these analyses, taxa abundances were fourth root transformed $n+0.1$ and the similarity coefficient was Bray-Curtis (PRIMER-E Ltd. 2001). "Bubble" plots were superimposed over the MDS projection representing selected taxa and physico-chemical variables to facilitate interpretation of the MDS analysis.

Numerical classification analysis (PRIMER-E Ltd. 2001) was used to investigate the structure of the benthic community (site x year and taxa). The site x year structure was evaluated using fourth root transformed n+0.1 abundances (all taxa). Biotic structure was evluated using RIMERhe 50 most abundant taxa (standardized densities). The similarity measure was Bray-Curtis and the clustering algorithm was "group average". PRIMER's SIMPER (PRIMER-E Ltd. 2001) program was used to rank the various taxa's contribution to the dissimilarity between identified clusters.

PRIMER's BIO-ENV (PRIMER-E Ltd. 2001) program was used to determine the association (weighted Spearman rank correlation) between the benthic community similarity matrix (fourth root transformed n+0.1 abundances; Bray-Curtis similarity) and 36 physical, hydrographic, and contaminant variables $(Log₁₀ (x+1)$ transformed and standardized; normalized Euclidean distance) for the 1995-1998 data* (Clarke & Ainsworth 1993).

Sediment type $(e.g., \text{ sand}, \text{silt})$ was determined by regressing %SC vs. mean grain (ϕ) size for Tampa Bay data collected by Long *et al*. (1994) using TableCurve 2D (AISN Software, 2000). These data were used to develop a relationship between %SC and mean grain size: %SC= $1/(0.0097+1.575*e$ ^{\$} ϕ [{]adjusted r²=0.947). Wentworth size classes for sediments (*cf*. Percival & Lindsay 1997) were then estimated for each %SC value.

* 1993 data were excluded from this analysis because only four samples were collected for chemical contaminants; 1994 samples were excluded because sediment contaminants were not analyzed.

RESULTS

Hydrographic: Table 1 summarizes the surface and bottom water quality measures, as well as sample depth, for the 139 stations sampled. Median sample depth was 2.4-m, although depths ranged to >14-m (Figure 2). The deepest stations (>11 m) were located in the East Bay and Pendola Point areas. The shallowest sites (0.1 m) tended to be located along the eastern shoreline between the Alafia River and Pendola Point.

Table 1. Summary of Mean Physicochemical Variables: Hillsborough Bay, 1993-1998

A. SURFACE

B. BOTTOM

Figure 2. CDF plot of sample depths in Hillsborough Bay, 1993-1998 inclusive.

The temperature-salinity plot suggests that the near-bottom water mass characteristics differed among years (Figure 3). Highest water temperatures were observed during 1995 and in 1993 (Figure 4). The frequency distribution of water temperatures during 1994 were similar to 1997 and 1998; 1996 was similar to 1997 (KS test $p > 0.05$). Salinities were generally highest in 1997 and lowest during the 1995 and 1998 sampling periods (Figure 5). The frequency distributions of near-bottom salinities were similar during 1993 and 1996 when polyhaline (>18 ppt) conditions predominated. Salinities were also similar during 1998 and 1995 when mesohaline (5-18 ppt) salinities were established in parts of Hillsborough Bay (KS test $p > 0.05$) (Figure 6). Rainfall data (Appendix B) for the area suggested that 1994 and 1995 were the wettest years and 1993 and 1998 were the driest during this study period.

Figure 3. Temperature-salinity plot: Hillsborough Bay 1993-1998. Ellipses embrace + 1 S.D. within each year.

Near-bottom dissolved oxygen concentrations were hypoxic $(\leq 2$ ppm) in $\geq 30\%$ of samples collected during 1998 and <10% of 1996 samples (Figure 7). Hypoxia was most often observed in the upper portions of Hillsborough Bay (Figure 8). The frequency distribution of near-bottom DO during 1995 differed (generally higher) from that of 1996-1998 (KS test *p<*0.05). (Figure 7).

Sediment Characteristics: Sandy sediments (<25.95 %SC) predominate in Hillsborough Bay (Figure 9). Muddy sediments are more widely distributed in the western parts of Hillsborough Bay; coarser sediments predominate in the southeastern quadrant (Figure 10).

Figure 4. CDF plot of near-bottom temperatures in Hillsborough Bay, by year 1993-1998.

Figure 5. CDF plot of near-bottom salinities in Hillsborough Bay, by year 1993-1998.

.

Figure 6. Near-bottom salinity zones in Hillsborough Bay, 1993-1998. Green= polyhaline (18-30 ppt); Yellow=Mesohaline (5-18 ppt); Orange=oligohaline (0.5-5 ppt).

Figure 7. CDF plot of dissolved oxygen concentration in Hillsborough Bay, 1993-1998, by year.

The apparent redox potential discontinuity layer (RPD) ranged from 0 to >100 mm (Figure 11), although in more than half of the samples no RPD was evident. RPD was negatively correlated with %SC (Figure 12) and positively correlated with DO (Figure 13).

Principal Components Analysis [PCA] of Hydrographic and Site Characteristics

PCA showed that the first two PCs explained >60% of the overall variation in Hillsborough Bay hydrography and site characteristics (Table 2). The highest loadings (Table 2-B) in PC1 were for depth, %SC and DO. Temperature had the highest loading in PC2.

Figure 8. Near-bottom dissolved oxygen concentrations in Hillsborough Bay, 1993-1998. Green: >4 **ppm; Yellow:>2 <4 ppm; Red:<2 ppm**

Figure 9. CDF plot of percent silt+clay in Hillsborough Bay sediments, 1993-1998 inclusive. Vertical lines demarcate sediment types: coarse sand (<1.70% SC), medium sand (1.70 to 4.51 %SC), fine sand (4.51 to 11.35 %SC), very fine sand (11.35 to 25.95 %SC), coarse silt (25.95 to 49.28 %SC), medium silt (42.98 to 89.98 %SC), and fine silt (>89.98 %SC).

The ordination plot and companion "bubble" plots (Figure 14) showed that the deeper sites had higher %SC and DO lower than shallower sites (Pincipal Component 1). Water temperature had the highest loading in Principal Component 2, reflecting interannual differences: cooler in 1998 than in 1993 and 1995 (Figure 14; *cf*. Figure 4).

Figure 10. Map depicting the distribution of sediment types in Hillsborough Bay, 1993- 1998.

Figure 11. CDF of apparent redox potential discontinuity layer [RPD] in Hillsborough Bay, 1993-1998. Anaerobic [AN] sediments are characterized by an RPD <10-mm and aerobic [AER] sediments are characterized by an RPD>50-mm.

Figure 12. Association between apparent RPD and %SC in Hillsborough Bay, 1993- 1998.

Figure 13. Association between apparent RPD and near-bottom dissolved oxygen, Hillsborough Bay, 1993-1998.

Table 2. Summary of PCA for hydrographic and site variables: Hillsborough Bay 1993-1998

A. EIGENVALUES & VARIANCE EXPLAINED

B. EIGENVECTORS

Figure 14. PCA of sample sites in Hillsborough Bay, 1993-1998 and "bubble" plots of %SC, depth, DO, and temperature superimposed on the samples: by year.

Sediment Contaminants: A composite PEL Quotient (Figure 15) suggested that, in any year surveyed to date, <4.8% of the of the sediment samples collected from Hillsborough Bay had a high likelihood of being toxic to aquatic life. Degraded sediments were only found in the upper parts of Hillsborough Bay: McKay Bay, East Bay, and in the Lower Hillsborough River (Figure 16). Frequency distributions of the composite PEL quotient (excluding 1993 when only three samples were collected) showed that only 1997 and 1996 had equivalent (*p*>0.05) distributions. For metals (Figure 17) and PAHs (Figure 18), the percentage of "degraded" samples also ranged from 0 to 4.8 and for PCBs (Figure 19) the percentage ranged from 0 to 3.7 in any year. Although organochlorine [OCL] pesticides were not included in the computation of the composite PEL quotient, as many as 10% of the samples exceeded the PEL for chlordane in any year (Figure 20) and none of the samples exceeded the PEL for total DDT (Figure 21).

Benthic Community: Table 3 summarizes selected benthic community measures for 1993-1998. At least 315 taxa were identified during this period (Appendix C). Numerically abundant species included the bivalve *Mysella planulata,* several polychaete worms (*Monticellina dorsobranchialis*, *Carazziella hobsonae, Prionospio perkinsi*), tubificid oligochaetes, and the amphipod crustacean *Ampelisca holmesi.*

Numerical dominants differed both by year (Table 4) and salinity zone (Table 5). The bivalve *M. planulata* was ranked either first or second during four of the six years, tubificid oligochaetes were ranked in the top ten during each year, and the polychaete *M. dorsobranchialis* was ranked among the ten dominants during five of the years. Three taxa were ranked in the top ten in both the oligohaline and mesohaline zones and four taxa were ranked in the mesohaline and polyhaline zones.

Figure 15. CDF plot of the composite (metals, PAHs, PCBs) PEL quotient for sediment contaminants in Hillsborough Bay, by year. Vertical lines demarcate "clean" (<0.05) and "degraded" (>0.34) sediments.

Figure 16. Composite PEL Quotients of Hillsborough Bay sediments, 1993 & 1995-1998. Green:<0.05; Yellow:>=0.05<0.34 ; Red:>0.34

Figure 17. CDF plot of the PEL quotient for metals (composite) in Hillsborough Bay, by year. Vertical lines demarcate "clean" (PEL quotient <0.1) and "degraded" (PEL quotient >1) sediments.

Figure 18. CDF plot of total PAH concentrations in Hillsborough Bay, by year. Vertical lines demarcate TEL (1684) and PEL (16770 ppb).

Figure 19. CDF plot of total PCB concentrations in Hillsborough Bay, by year. Vertical lines demarcate TEL (21.6 ppb) and PEL (189 ppb).

Figure 20. CDF plot of total chlordane concentrations in Hillsborough Bay, by year. Vertical lines demarcate TEL (2.26 ppb) and PEL (4.79 ppb).

Figure 21. CDF plot of total DDTs concentrations in Hillsborough Bay, by year. Vertical line demarcates the TEL (3.89 ppb).

	Abundance $(\frac{\text{H}}{m^2})$	Species Richness (S)	Diversity (H')	Evenness $\bf (J)$	TBBI
Minimum				0.00	-3.0
Maximum	35750	54	4.32	.00	26.1
Median	.225		2.53	0.58	
Mean				0.55	127

Table 3. Summary of Benthic Community Measures: Hillsborough Bay, 1993-1998

Numbers of taxa per station were variable over years (Figure 22). During 1995, the median numbers of taxa per m² was <20 whereas in 1994 it was >30. During 1995, almost 20% of the samples were devoid of living organisms. The KS test showed that the frequency distribution during 1993 and 1996 a greater proportion of samples had >20 taxa samples than during 1998. In 1995 proportionately more samples had <10 taxa than during 1997.
1993	1994	1995	1996	1997	1998			
1.Mediomastus ambiseta	1. M. planulata (599)	1. Amygdalum papyrium (541)	1. M. planulata (2319)	M. planulata (1422)	P. perkinsi (808)			
$(1499*)$ 2. Mysella planulata (947)	2. M. ambiseta (572)	2. A. holmesi (537)	2. M. dorsobranchialis (578)	A. holmesi (1283)	M_{\cdot} dorsobranchialis (392)			
3. Carazziella hobsonae (645)	3. Tubificidae- gen. undet. (571)	3. Tubificidae- gen. undet. (354)	3. C. hobsonae (458)	M_{\cdot} dorsobranchialis (681)	Paramphinome sp. B(252)			
4. Ampelisca holmesi (494)	4. M. dorsobranchialis (451)	4. Streblospio gynobranchiata (256)	4. A. holmesi (367)	R. naglei (605)	Pinnixa $sp(p)$. (207)			
5.Tubificidae- gen. undet. (403)	5. A. holmesi (255)	5. P. pinnata (208)	5. Laeonereis culveri (365)	P. perkinsi (351)	Tubificidae-gen. undet. (202)			
6. Branchiostoma floridae (355)	6. P. perkinsi (222)	6. Mulinia lateralis (195)	6. A. philbinae (275)	Cerapus sp. C (325)	C. hobsonae (197)			
\mathcal{I} . Paraprionospio pinnata (351)	7. P. pinnata (209)	7. M. planulata (188)	7. Rudilemboides naglei (226)	P. triquetra (318)	Enteropneusta- gen. undet. (171)			
8. Monticellina dorsobranchialis (327)	8. B. floridae (188)	8. Nereis succinea (162)	8. Tubificidae- gen. undet. (226)	Tubificoides wasselli (317)	P. pinnixa (133)			
9. Prionospio perkinsi (295)	9. Aricidea philbinae (183)	9. Parastarte triquetra (141)	9. Podarkeopsis laevifuscina (204)	C. hobsonae (297)	Glottidia pyramidata (131)			
10. Mediomastus sp.(p.) (278)	10. C. hobsonae (182)	10. Enchytraeidae- gen. undet.	10. B. floridae (174)	B. floridae (285)	Gyptis crypta (126)			

Table 4. Ten Most Abundant (mean # m-2) Macroinvertebrate Taxa in Hillsborough Bay, 1993-1998: By Year

Table 5. Ten Most Abundant (mean # m-2) Macroinvertebrate Taxa in Hillsborough Bay, 1993-1998: By Salinity Zone (Venice System)

OLIGOHALINE	MESOHALINE	POLYHALINE						
$(0.5-5.0 \text{ ppt})$	$(5.0-18.0 \text{ ppt})$	$(18.0 - 30.0 \text{ ppt})$						
1. Streblospio gynobranchiata	1. A. holmesi (493)	1. M. planulata (1179)						
$(2075*)$								
2. Tubificoides brownae	2. Amygdalum papyrium (266)	2. A. holmesi (523)						
(1500)								
3. Stenoninereis martini (825)	3. Mulinia lateralis (246)	3. M. dorsobranchialis						
		(491)						
4. Cyclaspis cf. varians (800)	4. S. gynobranchiata (235)	4. Mediomastus ambiseta						
		(369)						
5. Tubificidae-gen. undet.	5. Mysella planulata (182)	5. Prionospio perkinsi						
(600)		(365)						
6. Ampelisca holmesi (325)	6. Monticellina	6. Carazziella hobsonae						
	dorsobranchialis (168)	(362)						
7. Cerapus sp. $C(275)$	7. Parastarte triquetra (138)	7. Tubificidae-gen. undet.						
		(350)						
8. Cyathura polita (100)	8. Paraprionospio pinnata	8. Rudilemboides naglei						
	(100)	(231)						
9. Ampelisca sp. $C(75)$	9. Tubificidae-gen. undet. (88)	9. Branchiostoma floridae						
		(218)						
9. Hartmanodes nyei (75)	10. Aricidea philbinae (62)	10. P. pinnata (190)						
9. Ambidexter symmetricus								
(75)								

Tampa Bay Benthic Index (TBBI) scores were generally higher during 1993, 1995 and 1996 than during 1997 and 1998 (Figure 23); the KS test showed that frequency distributions differed for these two groups of years. TBBI scores <4.6 (degraded habitat) were found at approximately 30% of the sites sampled during 1995, contrasted with <15% of sites in other years. Degraded habitat was generally located immediately south of the Hillsborough River and in the northeastern industrial segment of Hillsborough Bay (Figure 24).

Correlation analysis showed that the TBBI was associated with DO $(r=0.54; p<.001)$ (Figure 25), %SC (r=-0.44; *p*<.001) (Figure 26), RPD (r=0.34; *p*<.001) (Figure 27), the composite PEL quotient (r=-0.31; *p*<.01) (Figure 28), and depth (r=-0.27; *p*<.001) (Figure 29), but not with salinity (r=0.04; *p>*.05) (Figure 30).

The stepwise multiple regression (adjusted multiple $r^2 = 0.33$; *p*<.001; n=120) was: TBBI $(log_{10} n+1) = 0.804 - 0.002*RPD(log_{10} n+1) + 0.524*DO(log_{10} n+1) - 0.239*%SC$ (ASN)

Figure 22. CDF for numbers of taxa in Hillsborough Bay benthos, by year, 1993-1998.

Figure 23. CDF of the Tampa Bay Benthic Index for Hillsborough Bay benthos, by year, 1993-1998. Index scores <4.6 indicate "degraded" benthic habitat and scores >4.6 indicate "healthy" benthic habitat.

Figure 24 Distribution of "healthy" (green) and "degraded" (red) benthic habitat in Hillsborough Bay, 1993-1998 based upon the Tampa Bay Benthic Index.

Figure 25. Association between the Tampa Bay Benthic Index and near-bottom dissolved oxygen concentrations in Hillsborough Bay, 1993-1998. TBBI scores <4.6 indicate "degraded" habitat and scores >4.6 indicate "healthy" habitat. DO <2.0 ppm indicates "degraded" conditions, >2<4 ppm indicates "marginal" conditions, and >4 ppm indicates "healthy" conditions.

Figure 26. Association between the Tampa Bay Benthic Index and %SC in Hillsborough Bay, 1993-1998. TBBI scores <4.6 indicates "degraded" habitat and scores >4.6 indicate "healthy" habitat. Sand and mud fractions are demarcated by 25.95%SC.

Figure 27. Association between the Tampa Bay Benthic Index and the apparent redox potential discontinuity layer (RPD) in Hillsborough Bay, 1993-1998. TBBI scores <4.6 indicate "degraded" habitat and scores >4.6 indicate "healthy" habitat.

Figure 28. Association between the Tampa Bay Benthic Index and the composite PEL quotient in Hillsborough Bay, 1993-1998. TBBI scores <4.6 and and scores >4.6 indicate "healthy" habitat. PEL quotients >0.34 indicate "degraded" sediments and PEL quotients <0.05 indicate "clean" sediments.

Figure 29. Association between the Tampa Bay Benthic Index and sample depth in Hillsborough Bay, 1993-1998. TBBI scores <4.6 indicate "degraded" habitat and scores >4.6 indicate "healthy" habitat.

Figure 30. Association between the Tampa Bay Benthic Index and salinity in Hillsborough Bay, 1993-1998. TBBI scores <4.6 indicate "degraded" habitat and scores >4.6 indicate **"healthy" habitat. Salinity zones are demarcated as: oligohaline (<5 ppt), mesohaline (5-18 ppt), and polyhaline (18-30 ppt).**

Benthic Community Structure: Two primary and four secondary "clusters" were identified in the classification analysis of sites (Figure 31) and 9 "clusters" were identified in the classification analysis of taxa (Figure 32). SIMPER analyses (Clarke & Warwick 2001) showed that dissimilarities between the biotic assemblages in Clusters A and B were primarily influenced by the higher densities of the bivalve mollusc *M. planulata*, tubificid oligochaete worms*,* the polychaete worm *P. perkinsi* and the amphipod crustacean *A. holmesi* in Cluster B (Table 6 and Appendix D). With the exception of *P. perkinsi*, these taxa were associated with an assemblage of species (Figure 32; Table 7) that also included the polychaetes *Mediomastus ambiseta, Aricidea philbinae*, and *Paraprionospio pinnata*.

These clusters also differed in their site characteristics, including hydrographic and sedimentary characteristics (Table 8). The assemblage characterized by Cluster A was considerably "poorer" biologically and subject to greater environmental stress than Cluster B. Cluster A sites had fewer taxa, a lower TBBI score and considerably lower densities of organisms. The A sites were deeper than the B sites and were characterized by a shallower RPD and higher % SC. The effects of stressors such as DO and sediment contaminants were greater at Cluster A sites as well.

Cluster A could be subdivided into clusters A-1 and A-2. The benthic assemblages at Cluster A-1 sites were especially depauperate (Tables 9 and 10). The stresses from low DO and sediment contaminants appeared to be greater at the A-1 sites as well (Table 10).

Cluster B could be further subdivided into clusters B-1 and B-2 as well. B-1 sites were generally located in shallow waters (Table 11) near the periphery of southern Hillsborough Bay and comprised samples collected during 1996 and 1997. The species contributing most to the dissimilarity between these two clusters included *M. planulata* and two amphipods (*A. holmesi* and *Rudilemboides naglei*), each of which were extremely abundant at B-1 stations relative to B-2 (Table 11). Other species which were especially abundant in B-1 sites included *Parastarte triquetra* (Bivalvia), *Laeonereis culveri* (Polychaeta), *Cerapus* sp. C (Amphipoda) and *Tubificoides wasselli* (Tubificidae) (Table 7). B-2 sites included species in Cluster 5 and Cluster 8 among its numerical dominants. Average total abundance was quite high in B-1 as was the mean numbers of taxa (Table 12). The mean TBBI, however, was higher in B-2-- a consequence of higher diversity and evenness at the B-2 sites. Of the measured stressors, chlordane concentrations were higher at B-2 sites and metals were higher at the B-1 sites.

Table 6. Results of SIMPER analysis comparing the dissimilarity between Hillsborough Bay Clusters A and B. Average dissimilarity = 15.15.

Linking Biotic & Abiotic Variables (1995-1998): PRIMER's (Primer-E Ltd. 2001) BIO-ENV procedure was used to explore the extent to which the benthic community structure can be explained by the measured physico-chemical characteristics. In order to maximize the physicochemical variable list (site characteristics, DO, chlordane, DDT, PCBs, total PAHs, and metals), the analysis was restricted to 1995-1998 data. The rank correlations between the biological data (Figure 33) and the "best fit" for physico-chemical variables (chlordane, PCBs, Al, Cu) (Figure 34) were very weak $(0.08).$

Interannual Trends: MDS of average abundance of all benthic taxa, by year, show that the benthic community experienced more pronounced shifts in structure from 1996 to 1997 and from 1997 to 1998 (Figure 35). BIO-ENV analysis using site, hydrographic, and contaminant (metals only; no organic data for 1994) variables showed that the best fit for environmental variables with the biotic data were for As, Cd, and Cr (Spearman r=0.84). As, Cd and Cr each showed a marked increase in mean concentration from 1996 to 1997 (Figure 35). From 1997 to 1998 As concentrations declined approximately 75%and Cr increased almost five fold.

cluster

Figure 32. Dendrogram depicting the similarity of 50 most abundant taxa (standardized abundance; Bray-Curtis similarity; Group-average clustering): Hillsborough Bay 1993-1998.

TAXA CLUSTER TAXA		$A-1$	$A-2$	$B-1$	$B-2$ a1	$B-2$ a2	$B-2$ $b1$	$B-2$ $\mathbf{b}2$
1	ENCHYTRAEIDAE							137
$\overline{2}$	ECHINOIDEA		5	150			1	
	Synaptidae A			497	$\overline{3}$			\overline{c}
3	Kalliapseudes sp. A		$\overline{4}$					$\mathbf{1}$
4	Teinostoma sp.		14	9		$\overline{2}$	39	22
5	Monticellina dorsobranchialis	$\overline{2}$	20	53	132	2658	53	735
	OPHIUROIDEA			66	34	96	26	78
	Hemipholis elongata		$\overline{4}$	297	6	179	14	50
	Ancistrosyllis jonesi		S			23	3	5
	Cerebratulus lacteus					56	23	$\mathbf{1}$
	Pinnixa spp.		$\overline{7}$	38	29	375	89	94
	Prionospio perkinsi		13	200	237	1802	108	290
	Sigambra tentaculata		11	103	19	194	61	97
	Paramphinome B	$\overline{2}$	14	31	8	350	55	53
	Gyptis crypta		11	19		231	36	32
	ENTEROPNEUSTA		32	34	\mathfrak{S}	350	18	111
	Sigambra bassi		8			27	22	8
	Carazziella hobsonae		21	34	88	1100	191	623
6	Glottidia pyramidata			53	$\overline{4}$	21	82	23
	Haminoea succinea		1	256	233		26	74
	Rudilemboides naglei		1	2078	369	$\overline{2}$	$\overline{3}$	$\overline{\mathbf{3}}$
	Tubificoides wasselli			856			29	42
	Cerapus sp. C (="tubularis")			863	1			11

Table 7. Two-way coincidence table (taxa by cluster), Hillsborough Bay benthos, 1993- 1998.

TAXA CLUSTER TAXA		$A-1$	$A-2$	$B-1$	$B-2$ a1	$B-2$ a2	$B-2$ $b1$	$B-2$ $\mathbf{b}2$
7	Parahesione luteola		8		12	8	14	41
	Streblospio gynobranchiata	$\overline{2}$	1	3	239	29	51	303
	Aricidea philbinae		1	72	635	13	26	18
	Scolelepis texana			41	84		12	
	Apoprionospio pygmaea	$\overline{2}$		34	16	19	$\overline{2}$	6
	Podarkeopsis levifuscina		11	313	161	144	33	74
	Mysella planulata		5	7113	1905	69	81	392
	Mediomastus ambiseta			3	1637	263	18	107
	Paraprionospio pinnata	8	18	34	292	292	181	224
	TUBIFICIDAE	$\overline{4}$	31	253	390	508	241	637
	Mulinia lateralis			78	119	98	126	120
	Ampelisca abdita		45	37.6	68	25	11	11
	Pectinaria gouldii	$\overline{2}$		9	92		14	37.6
	Amygdalum papyrium		$\overline{4}$	269	813		62	78
	Nereis succinea		$\overline{4}$	106	288		17	129
	Ampelisca holmesi		51	3353	1259	8	182	224
8	Mediomastus sp.		1	16	278	140	9	25
	Mediomastus californiensis			9	6	29	62	6
	Caecum strigosum		$\overline{4}$			669	23	15
	Branchiostoma floridae			263	153	792	168	71
9	Parastarte triquetra			1178	283		$\overline{2}$	3
	Bittiolum varium			519	6			
	Laeonereis culveri		11	875	230	$\overline{4}$	11	120
	Cyathura polita			59	29		10	10
	Phoronis sp.		$\mathbf{1}$	$\overline{3}$	1		8	65
	Stenoninereis martini		15				2	61
	Tubificoides brownae		$\mathbf{1}$	9	13	$\overline{4}$	15	132

Table 7 (continued). Two-way coincidence table (taxa by cluster), Hillsborough Bay benthos, 1993-1998.

Table 8. Comparison of mean site characteristics, hydrographic conditions, sedimentary contaminants, and biotic variables: Cluster A *vs***. Cluster B (***cf***. Figure 30), Hillsborough Bay, 1993-1998.**

Table 9. Results of SIMPER analysis comparing the dissimilarity between Hillsborough Bay Clusters A-1 and A-2. Average dissimilarity = 3.03.

Table 10. Comparison of mean site characteristics, hydrographic conditions, sedimentary contaminants, and biotic variables: Cluster A-1 *vs***. Cluster A-2 (***cf***. Figure 30), Hillsborough Bay, 1993-1998.**

Table 11. Results of SIMPER analysis comparing the dissimilarity between Hillsborough Bay Clusters B-1 and B-2. Average dissimilarity = 24.92.

Table 12. Comparison of mean site characteristics, hydrographic conditions, sedimentary contaminants, and biotic variables: Cluster B-1 *vs***. Cluster B-2 (***cf***. Figure 30), Hillsborough Bay, 1993-1998.**

Figure 33. Non-Metric Multi-Dimensional Scaling (MDS) representation of benthic community structure in Hillsborough Bay, 1995-1998, by year

Figure 34. "Bubble" plots of chlordane, PCBs, Al, and Cu concentrations superimposed over the MDS plot depicting benthic community structure, by year, in Hillsborough Bay 1995-1998

Figure 35. MDS plot of "average" benthic community structure by year and mean concentrations of As, Cd, and Cr by year, Hillsborough Bay, 1993-1998. Lines delineate temporal trend.

Shifts in benthic community structure from 1996 to 1997 were primarily affected by declines in the abundance of *M. planulata* and increases in the abundances of *Prionospio perkinsi* and *A. holmesi* (Appendix E; Figure 36). Taxa which contributed to the shift in benthic community structure from 1997 to 1998 include *M. planulata*, *M. dorsobranchialis*, and *A.holmesi,* each of which underwent a large decline in abundance (Appendix E; Figure 36).

Status of Hillsborough Bay Sediments: Hillsborough Bay appears to be affected more by subnominal DO than the Louisianian Province (northern Gulf of Mexico south to Tampa Bay) as a whole (Table 13). Degradation of benthic habitat, however, appears to be markedly higher in the Louisianian Province than in Hillsborough Bay.

Figure 36. Mean densities (# m-2) concentrations of *Mysella planulata, Ampelisca holmesi***, tubificid oligochaetes (genera undetermined),** *Monticellina dorsobranchialis***, and** *Prionospio perkinsi***, in Hillsborough Bay, 1993-1998.**

Table 13. Comparison of proportions of degraded habitat, by category and study area: Hillsborough Bay, Tampa Bay, Florida (1993-1998) (as % of samples) *vs***. Louisianian Province (as % area)**.

^a Summers *et al.* 1993 ^b. Macauley *et al.* 1994 ^c Macauley *et al.* 1995

DISCUSSION

Hillsborough Bay is the most industrialized segment of Tampa Bay as well as one of the most urban segments. Hillsborough Bay's sediments are exposed to nutrients and other contaminant inputs from urban stormwater, industrial and municipal wastewater discharges, thermal discharges and atmospheric deposition from power plants, and the phosphate industry (Estevez 1989). These impacts have been expressed as pervasive algal blooms (FWPCA 1969), subnominal DO, faunal die-offs (Santos & Simon 1980), and contamination of sediments by metals, pesticides, and PAHs (Doyle *et al.* 1989; Long *et al*. 1994; Grabe 1997; Grabe & Barron 2002). Improved treatment of municipal and industrial wastes, increased treatment, retention, and detention of stormwater, and reductions in industrial emissions have all served to ameliorate these problems.

During 1993-1998, 23% of the Hillsborough Bay samples met at least one criterion for "degraded" benthic habitat. DO was subnominal (<2 ppm) at 17%of the sites-- including approximately 30% of the 1995 samples. These areas were generally located downstream of the bay's confluence with the Hillsborough and Alafia rivers. Coincidentally, during 1995 June- August rainfall at sites along the Lower Hillsborough (SWFWMD site 376) and Alafia (SWFWMD site 252) rivers were higher than any year except 1994 and that during 1995 more shallow, inshore sites were sampled than during the other years. The association between rainfall, runoff and stream flow and the dissolved oxygen status of Hillsborough Bay merits more detailed examination.

Based upon the TBBI cutoff of 4.6 (MacDonald *et al*. 2002), the benthic assemblages at 10% of the siteswere also degraded. Degraded benthic assemblages were most often located immediately bayward of the mouth of the Hillsborough River during 1995; a secondary area of degraded benthos was the industrial area of northeastern Hillsborough Bay.

Almost 7% of the samples were subnominal for both DO and benthic structure. Only 3% of the samples were subnominal with respect to sediment contaminants based upon a composite PEL quotient. For metals and PAHs the percentage of "degraded" samples also ranged from 0% to <5% and for PCBs the percentage ranged from 0% to 3.7% in any year.

As many as 10% of the samples exceeded the PEL for chlordane in any year and none of the samples exceeded the PEL for total DDT. The portions of Hillsborough Bay with sediments degraded by chemical contaminants were located in McKay Bay, East Bay, and the lower Hillsborough River.

Correlation analysis showed that the TBBI was positively associated with DO and RPD and negatively associated with %SC, the composite PEL quotient, and depth, but not with salinity. The rank correlations between the biological data and the "best fit" for physicochemical variables and chemical contaminants (chlordane, PCBs, Al, Cu) were very weak $(**0.1**).$

During the study period, the near-bottom water mass characteristics differed among years. Near-bottom salinities in Hillsborough Bay were generally in the polyhaline (18-30 ppt) zone. Both mesohaline and oligohaline salinities were observed, especially during 1995 and 1998. Salinity patterns did not show any consistent association with rainfall patterns in the near-field. Highest total rainfall during June through August occurred during 1994 (Appendix B) when very little of Hillsborough Bay had salinities <18 ppt. Mesohaline salinities were most prevalent during 1998 when combined rainfall totals for SWFWMD stations in the Lower Hillsborough and Alafia rivers were fourth highest over the six year period (Appendix B). Other variables, such as streamflow, stormwater runoff, and rainfall at other locations in the watershed should be examined to determine their association with salinities in Hillsborough Bay. The averages reported in this survey period (20.1 ppt surface/21.6 bottom) are consistent with the long-term mean (20.9 pb) reported by Simon (1974).

PCA showed that depth, %SC and DO exerted primary influence on the intra-bay habitat characteristics; salinity was only a minor contributor. %SC was higher and DO was lower at the deeper sites. Temperature exerted a secondary effect, apparently as a surrogate for interannual trends, rather than as an indicator of intra-bay spatial patterns. Sandy sediments predominate in Hillsborough Bay with the coarsest sediments in the southeastern quadrant. Muddy sediments are widely distributed in the western parts of Hillsborough Bay. Doyle *et al*. (1989) and Johansson & Squires (1989), using somewhat different criteria for "mud" *vs*.

60

"sand", described a similar distribution using data collected during the 1960s. and 1980s. Although the historical data are sparser than the current data, the location and extents of fine-grained sediments in Hillsborough Bay seems to have remained fairly consistent over the past 40 years.

Within the benthic community, numerically abundant species included *Mysella planulata* (Bivalvia), several polychaete worms (*Monticellina dorsobranchialis*, *Carazziella hobsonae, Prionospio perkinsi*), tubificid oligochaetes, and *Ampelisca holmesi* (Amphipoda). Numerical dominants differed both by year and salinity zone. *Mysella planulata* was ranked either first or second during four of the six years, tubificid oligochaetes were ranked in the top ten during each year, and *M. dorsobranchialis* was ranked during five of the years. Three taxa were ranked in the top ten in both the oligohaline and mesohaline zones and four taxa were ranked in the mesohaline and polyhaline zones.

Data provided by Taylor *et al*. (1970), Taylor (1971), Thoemke (1979), and Santos & Bloom (1980) provide an opportunity to compare the species composition of Hillsborough Bay benthos between 1963-1964, 1975-1977, and 1993-1998. The most frequently occurring mollusks reported by Taylor *et al.* (1970) included *Mulinia lateralis*, *Nassarius vibex*, *Amygdalum papyrium, Tellina versicolor*, and *Macoma tenta*. These were all among the ten most frequently occurring mollusks during the current survey period as well. *Mysella planulata* occurred in >48% of the 1993-1998 samples although it only occurred in 4% of the 1963-1964 samples. *Mysella planulata* was also among the dominants in Hillsborough Bay during the late 1970s (Santos & Simon 1980). *Prunum apicinum* was found in 30% of the samples collected during 1993-1998 but was absent from the 1963-1964 samples.

The most frequently collected polychaetes in Hillsborough Bay during 1963-1964 also differed from the current study. Taylor (1971) found *Spiochaetopterus costarum, Streblospio benedicti* (=*gynobranchiata*) and *Paraprionospio pinnata* in >60% of the samples. In the current database, the common polychaetes included *P. perkinsi* (47% of samples), *P. pinnata* (44%), and *C. hobsonae* (42%). Of the top five most frequently occurring polychaetes in this study only *P. pinnata* was similarly ranked in 1963. *Mediomastus* spp., which was a dominant in 1993 and were also dominants in the late 1970s (Santos & Bloom 1980), were not reported by Taylor (1971) for any sample in Tampa Bay during 1963-1964.

Uebelacker *et al*. (1984) report that *M. californiensis* ranges up to 25 mm in length by 0.5 mm in width and Hartman (1947) reports that the dimensions of *M. ambiseta* are up to 15 mm by 0.5 mm. Therefore, even though sieve sizes were larger in the 1963 survey (0.7 mm; Taylor 1971) than in the current survey, *Mediomastus* spp. are large enough that some should have been retained by the 0.7 mm were they present in Taylor's samples.

Another apparent difference in species composition since the 1970s to the present is the change in *Ampelisca* spp. populations. Thoemke (1979) described the life history paprameters of *A. abdita* and *A. verrilli* in Hillsborough Bay during 1975-1976. *Ampelisca abdita* was quite abundant relative to *A. "verrilli*". Santos & Bloom (1980) also include these two species in their species list and include *A. abdita* among the numerical dominants. In the present study *A. holmesi* is among the numerical dominants in most years and *A. abdita* is less abundant. *Ampelisca holmesi* was described, by Pearse (1908), from specimens collected in Oyster Bay (near the St. Marks River in northern Florida), with subsequent reports of its occurrence in Charlotte Harbor (Pearse 1912) and Sarasota Bay (Shoemaker 1933. It seems likely that Thoemke's *A. "verrilli"* may be *A. holmesi*. That being the case, the dominance of *A. holmesi* in Hillsborough Bay during the current study period does appear somewhat different from data collected by Thoemke (1979) and Santos & Simon (1980) during the 1970s. However, this may merely reflect interannual trends in the populations of these species.

The benthic community of Hillsborough Bay apparently has changed since the 1960's-1970's--even taking into account interannual variation in composition and abundance.

Spatial and temporal dissimilarities between the biotic assemblages reported in this study in were primarily influenced by differences in the distributions of *M. planulata*, tubificids*, P. perkinsi and A. holmesi.* One of the two primary assemblages was considerably "poorer" biologically, perhaps an effect of greater environmental stress than the second primary cluster. Cluster "A" sites had fewer taxa, a lower TBBI score and considerably lower

62

densities of organisms; "A" sites were deeper than the "B" sites and were characterized by a shallower RPD and higher % SC. The effects of stressors such as DO and sediment contaminants were greater at Cluster "A" sites as well

The apparent RPD width demarcates reduced and oxidized sediments (Rosenberg 2001). The depth of the upper, oxidized layer is influenced by bioturbation (Rosenberg 2001;Rosenberg *et al*. 2001). In order for bioturbation to occur, the near-bottom DO regime must be adequate to sustain a diverse benthic assemblage (Nilsson $\&$ Rosenberg 2000). Thus, the width of the RPD and DO are correlated (Rosenberg 2001). RPD was, in fact, positively correlated with DO and negatively correlated with %SC in Hillsborough Bay.

Summers *et al*. (1993) suggested that, for Louisianian Province estuaries, an RPD <10mm may be indicative of anaerobic sediments and an RPD>50 mm may represent aerobic sediments. Using these criteria, Hillsborough Bay sediments were considerably more anaerobic than those of the Louisianian Province as a whole. In >50% of the Hillsborough Bay samples the RPD was not detected and in approximately 60% of the Hillsborough Bay samples the RPD was <10 mm, indicative of anaerobic sediments. This compares with 9% of the sediments in the estuaries of the Louisianian Province in 1991 (Summers *et al.*1993) At the Hillsborough Bay sites, however, both DO and %SC varied widely. Rosenberg *et al*. (2001) cautioned that the RPD may vary several centimeters over a short distance, even in hypoxic sediments.

Cluster "A" could be subdivided into two clusters. The benthic assemblages in Cluster "A-1" sites were especially depauperate and stresses from low DO and sediment contaminants were apparently greater at these sites.

Cluster "B" could also be subdivided. "B-1" sites were generally located in shallow waters near the periphery of southern Hillsborough Bay during 1996 and 1997 only. The species contributing most to the dissimilarity between these clusters "B-1" and "B-2" included *M. planulata* and two amphipods (*A. holmesi* and *Rudilemboides naglei*), each of which were extremely abundant at "B-1" stations. The mean TBBI, however, was higher in "B-2"-- a consequence of higher diversity and evenness at the "B-2" sites. Of the measured stressors, chlordane concentrations were higher at "B-2" sites and metals were higher at the "B-1" sites.

The benthic community experienced more pronounced shifts in structure from 1997 to 1997 and from 1997 to 1998 than during other sequential years. The best fit for the *mean* environmental variable data with the *mean* biotic data were for As, Cd, and Cr (Spearman r=0.84). As, Cd, and Cr each showed a marked increase in mean concentration from 1996 to 1997. From 1997 to 1998 As concentrations declined approximately 75% and Cr increased almost 5 fold. Taxa which primarily contributed to the shift in community structure from 1996 to 1997 included *M. planulata*, *M. dorsobranchialis*, and *A.holmesi*. Shifts in structure from 1997 to 1998 were primarily affected by changes in the abundance of *M. dorsobranchialis*, *Prionospio perkinsi*, and *A. holmesi.*

CONCLUSIONS

Soft-sediment habitats in portions of Hillsborough Bay experienced stress from low DO, trace metals, and chlordane. Low DO and subnominal benthic assemblages, were the two primary indicators of degraded benthic habitat. Low DO was more pervasive in Hillsborough Bay than in the Louisianian Province as a whole and subnominal benthic habitat was less pervasive in Hillsborough Bay.

The structure of the benthic community was not clearly explained by patterns or trends in physical factors generally shown to be linked to benthic community structure (*e.g*., salinity and sediment type). The linkage between biotic and abiotic structure was generally weak except when yearly survey averages were considered. In the interannual trend analysis, changes in the *average* concentrations of the metals As, Cd, and Cr were linked to changes in the *average* composition of the benthic community. The TBBI was, however, associated with several of the measured variables, including DO, RPD, depth, %SC, and a composite index of sediment contaminants.

Analysis of hydrographic (temperature, salinity) and habitat variables (depth, %SC, DO) suggested that sample depth, DO and %SC were primary determinants of the physicochemical "structure" of Hillsborough Bay. Water temperature, as an indicator of year-toyear changes, was a secondary factor. Salinity was less important in characterizing Hillsborough Bay.

The composition of the benthos appears to have undergone changes since the 1960s and 1970s. Changes were observed in the most frequently occurring species within three taxonomic groups: polychaete worms, mollusks, and ampeliscid amphipods. Although interannual variations in population size andlocation could explain some of these differences, the differences could indicate changes in habitat quality over the past 30 years.

We believe that the historical databases should be more rigorously analyzed in concert with the contemporary data (*cf*. Karlen et al. 1997). Such an approach may shed light on the

65

extent to which the benthic assemblages of Hillsborough Bay have changed concomitant with the documented improvements in water quality.

REFERENCES

AISN Software: 2000. *Table Curve 2D ver.5.0*, SPSS, Chicago, IL.

Clarke, K.R. & M. Ainsworth. 1993. A method of linking multivariate community structure to environmental variables*. Mar. Ecol Progr. Ser*. 92:205-219.

Clarke, K.R. & R.M. Warwick. 2001. *Change in Marine Communities: An Approach to Statistical Analysis and Interpretation.* 2nd Ed. PRIMER-E Ltd. Plymouth Marine Laboratory, UK.

Courtney, C.M., R. Brown, & D. Heimbuch. 1993*. Environmental Monitoring and Assessment Program Estuaries-West Indian Province: Volume I. Introduction, Methods and Materials, and Quality Assurance Field and Laboratory Operations Manual for a Synoptic Survey of Benthic Macroinvertebrates of the Tampa Bay Estuaries*. Environmental Protection Commission of Hillsborough County, Tampa, FL

Courtney, C.M., S.A. Grabe, D.J. Karlen, R. Brown, & D. Heimbuch. 1995*. Laboratory Operations Manual for a Synoptic Survey of Benthic Macroinvertebrates of the Tampa Bay Estuaries*. Environmental Protection Commission of Hillsborough County, Tampa, FL. [DRAFT]

Doyle, L.J., G.R. Brooks, K.A. Fanning, E.S. Van Vleet, R.H. Byrnes & N.J. Blake. 1989. *A Characterization of Tampa Bay Sediments*. Center for Nearshore Marine Science. USF. St. Petersburg. 99p.

ESRI, Inc. 1999. *ARCVIEW®* ver. 3.2. Redlands, CA.

Federal Water Pollution Control Authority (FWPCA). 1969*. Problems and Management of Water Quality in Hillsborough Bay, Florida.* Washington, DC. 86 p.

Grabe, S.A. 1997*. Trace Metal Status of Tampa Bay Sediments*. EPCHC, Tampa.

Grabe, S.A. & J. Barron. 2002. *Status of Tampa Bay Sediments Polycyclic Aromatic Hydrocarbons,Organochlorine Pesticides and Polychlorinated Biphenyls (1993 & 1995- 1999)*. EPCHC, Tampa.

Hartman, O. 1947. *Polychaetous annelids. Part VII. Capitellidae*. Allan Hancock Pacific Expeditions 10. USC Press. Los Angeles.

Karlen,D.J., S.Grabe, T.Perkins,W. Lyons, & G. Blanchard.1997. Tampa Bay benthos: Species composition of mollusks and polychaetes, 1963 *vs*.1993-1995. Pages 59-74. *In*: S.F.Treat (ed.). *Tampa Bay ScientificInformation Symposium 3. October 21-23, 1996.* Clearwater, FL. TEXT. St. Petersburg. 396 p.

REFERENCES (continued)

Long, E.R., D.A. Wolfe, R.S. Carr, K.J. Scott, G.B. Thursby, H.L. Windom, R. Lee, F.D. Calder, G.M. Sloane, & T. Seal. 1994. *Magnitude and Extent of Sediment Toxicity in Tampa Bay, Florida*. NOAA Tech. Mem. NOS ORCA 78.

Macauley, J.M., J.K. Summers, V.D. Engle, P.T. Heitmuller, G.T. Brooks, & M. Babikow. 1994*. Statistical Summary: EMAP-Estuaries Louisianian Province- 1992*. EPA/620/R-94/002.

Macauley, J.M., J.K. Summers, V.D. Engle, P.T. Heitmuller, & A.M. Adams. 1995*. Statistical Summary: EMAP-Estuaries Louisianian Province- 1993*. EPA/620/R-96/003.

MacDonald, D.D., R. A. Lindskoog, D.E. Smorong, H. Greening, R. Pribble, T. Janicki, S. Janicki, S. Grabe, G. Sloane, C.G. Ingersoll, D. Eckenrod, & E.R. Long. 2002 *Development of an Ecosystem-based Framework for Assessing and Managing Sediment Quality Conditions in Tampa Bay, Florida.* Prep. for Tampa Bay Estuary Program, St. Petersburg. [DRAFT].

Nilsson, H.C. & R. Rosenberg. 2000. succession in marine benthic habitats and fauna in response to oxygen deficiency:analyzed by sediment profile imaging and by grab samples. *Mar. Ecol Progr. Ser*. 197:139-149.

Pearse, A.S. 1908. Descriptions of four new species of amphipodous Crustacea from the Gulf of Mexico. *Proc. USNM* 34(1594):27-32.

Pearse, A.S. 1912. Notes on certain amphipods from the Gulf of Mexico, with descriptions of new genera and new species*. Proc. USNM* 43(1936):369-379.

Percival, J. B. & P.J. Lindsay. 1997. Measurement of physical properties of sediments. Pages 7-46. In A. Mudroch, J. M. Azcue and P. Mudroch (eds). *Manual of Physico-chemical Analysis of Aquatic Sediments.* Lewis Publ. Boca Raton.

PISCES Conservation Ltd. 2001. *Species Diversity and Richness II*. Lymington, England.

PRIMER. 2001. PRIMER-E, Ltd. Plymouth Marine Laboratory, UK.

Remane, A. 1934. Die brackwasserfauna. *Zool. Anz*., Suppl. 7:34-74.

Rosenberg, R. 2001. Marine Benthic faunal successional stages and related sedimentary activity*. Sci. Mar*. 65:107-119.

Rosenberg, R., H.C. Nilsson & R.J. Diaz. 2001. Response of Benthic fauna and changing sediment redox profiles over a hypoxic gradient. *Est. Coastal Shelf Sci.* 53:343-350.

Santos, S.L. & J.L. Simon. 1980. Response of soft-bottom benthos to annual catastrophic disturbance in a south Florida estuary. *Mar. Ecol Progr. Ser*. 3:347-3565.

REFERENCES (continued)

Shoemaker, C.R. 1933. Amphipoda from Florida and the West Indies. *Amer. Mus. Novit*. 598: 24 pp.

Simon, J.L. 1974. Tampa Bay estuarine system—a synopsis. *Fla. Sci*. 37:217-244.

SPSS Inc. 2000*. SYSTAT 10*. Chicago, IL.

Summers, J.K., J.M.Macauley, P.T. Heitmuller, V.D. Engle, A.M. Adams, & G.T. Brooks. 1993*. Statistical Summary: EMAP-Estuaries Louisianian Province- 1991*. EPA/620/R-93/007

Tampa Bay National Estuary Program. 1996. *Charting the Course: The Comprehensive Conservation and Management Plan for Tampa Bay*. St. Petersburg.

Taylor, J.L., J.R Hall & C.H. Saloman. 1970. Mollusks and benthic environments in Hillsborough Bay, Florida. *Fish. Bull*. 68:191-202.

Uebelacker, J.M., P.G. Johnson, & B.A. Vittor. 1984. *Taxonomic Guide to the Polychaetes of the Northen Gulf of Mexico. Volume 2. Barry Vittor & Assoc. Mobile, AL.*

Valente, R.M., D.C. Rhoads, J.D. Germano & V.J. Cabelli. 1992. Mapping of benthic enrichment patterns in Narragansett Bay, RI*. Estuaries* 15:1-17.

APPENDIX A HILLSBOROUGH BAY SAMPLING LOCATIONS: BY YEAR

APPENDIX B. MONTHLY RAINFALL (INCHES) AT THREE RAIN GAUGES OPERATED BY THE SOUTHWEST FLORIDA WATER MANAGEMENT DISTRICT, JUNE THROUGH AUGUST 1993-1998.

APPENDIX C INVENTORY OF BENTHIC MACROINVERTEBRATES COLLECTED FROM HILLSBOROUGH BAY, 1993-1998

Phylum Cnidaria **Class Anthozoa Order Actinaria** Actiniaria sp. A Actiniaria sp. B

Tribe Thenaria Family Actinostolidae Thenaria A

Phylum Platyhelminthes

Class Turbellaria Order Polycladida Turbellaria A *Eustylochus meridianalis*

Phylum Nemertea

Nemertea X Nemertea U Nemertea T Nemertea O Nemertea Q Nemertea N Nemertea P Nemertea L Nemertea G Nemertea F Nemertea E Nemertea I Nemertea K Nemertea B Nemertea A Nemertea J

Class Anopla Order Paleonemertea Family Tubulanidae *Tubulanus pellucidus Tubulanus sp. B*

Order Heteronemertea Family Celebratulidae *Cerebratulus lacteus*

Order Heteronemertea Family Lineidae *Zygeupolia cf. rubens?*

Class Enopla Order Haplonemertea Family Amphiporidae *Amphiporus bioculatus*

Phylum Annelida **Class Polychaeta Family Polynoidae** Polynoidae A *Malmgreniella maccraryae Malmgreniella taylori*

Family Sigalionidae Sthenelais A Family Crysopetalidae *Bhawania heteroseta*

Family Amphinomidae *Paramphinome B*

Family Phyllodocidae

Phyllodoce longipes Eteone heteropoda Eteone foliasa Nereiphylla castanea Paranaitis gardineri Nereiphylla fragilis Nereiphylla A Phyllodoce arenae

Family Hesionidae

Gyptis crypta Parahesione luteola Ophiodromus obscura Podarkeopsis levifuscina

Family Pilargidae

Ancistrosyllis sp. Ancistrosyllis hartmanae Ancistrosyllis jonesi Sigambra tentaculata Sigambra bassi Cabira incerta Synelmis ewingi Litocorsa antennata Litocorsa A

Family Syllidae

Pionosyllis sp. Syllis cornuta Exogone dispar Exogone arenosa Sphaerosyllis sp. Sphaerosyllis taylori Sphaerosyllis longicauda Sphaerosyllis labyrinthophila Brania wellfleetensis Brania A Parapionosyllis longicirrata Parapionosyllis uelebackerae

Family Nereididae

Nereis succinea Nereis falsa Nereis micromma Nereis lamellosa Laeonereis culveri Stenoninereis martini

Family Nephtyidae

Nephtys cf. hombergii Nephtys picta Nephtys cryptomma Aglaophamus verrilli

Family Glyceridae

Glycera Americana

Family Goniadidae

Glycinde solitaria Goniadides carolinae

Family Onuphidae

Onuphis A Diopatra cuprea Mooreonuphis cf. nebulosa Kinbergonuphis simony

Family Oenonidae

Arabella mutans

Family Dorvilleidae

Dorvillea rudolphi Pettiboneia sp.

Family Orbinidae

Leitoscoloplos robustus Scoloplos rubra Scoloplos texana Leitoscoloplos fragilis

Family Paraonidae

Aricidea suecica Aricidea cf. catherinae Aricidea philbinae Aricidea taylori Paraonis fulgens Cirrophorus sp. Paradoneis cf. lyra

Family Spionidae

Laonice cirrata Dipolydora socialis Polydora cornuta Prionospio heterobranchia Prionospio steenstrupi Apoprionospio pygmaea Prionospio cristata Apoprionospio dayi Prionospio perkinsi Spio pettiboneae Spio limnicola Paraprionospio pinnata Streblospio spp. Scolelepis texana Aonides mayaguezensis Carazziella hobsonae

Family Magelonidae

Magelona pettiboneae

Family Poecilochaetidae *Poecilochaetus johnsoni*

Family Chaetopteridae

Spiochaetopterus costarum

Family Cirratulidae

Caulleriella zetlandica Caulleriella D Tharyx sp. Monticellina dorsobranchialis Cirriformia A

Family Flabelligeridae *Piromis roberti*

Family Ophelidae

Armandia maculata Travisia hobsonae Polyopthalmus sp.

Family Capitellidae

Capitella capitata Capitella jonesi (=sp. A) Heteromastus filiformis Notomastus hemipodus Notomastus americanus Notomastus n. sp.? Mediomastus ambiseta Mediomastus californiensis Capitella jonesi

Family Maldanidae

Sabaco americanus Axiothella mucosa Axiothella A

Family Pectinariidae *Pectinaria gouldii*

Family Ampharetidae *Hobsonia florida Melinna maculata*

Family Terebellidae

Loimia sp.

Family Sabellidae

Megalomma pigmentum Fabricinuda trilobata

Family Polygordiidae

Polygordius sp.

Class Oligochaeta Order Tubificida Family Tubificidae *Limnodriloides monothecus complex*

Tubificoides brownae Tubificoides wasselli Thalassodrilides gurwitschi

Family Enchytraeidae

Phylum Mollusca **Class Gastropoda Family Hydrobiidae**

Family Vitrinellidae

Vitrinellidae sp. A Vitrinella floridana Teinostoma biscaynense Teinostoma megastoma

Family Caecidae

Caecum pulchellum Caecum strigosum

Order Mytiloida

Family Mytilidae *Modulus modulus*

Family Cerithiidae *Bittiolum varium*

Family Epitoniidae

Epitonium angulatum

Family Eulimidae

Melanella cf. arcuata Eulima bilineatus Microeulima hemphilli

Family Naticidae

Tectonatica pusilla

Family Columbellidae *Astyris lunata*

Family Melongenidae *Melongena corona*

Family Nassaridae *Nassarius vibex*

Family Olividae

Jaspidella blanesi Olivella floralia Olivella pusilla Oliva sayana

Family Marginellidae

Prunum apicinum

Family Conidae

Kurtziella limonitella

Family Pyramidellidae

Fargoa cf. gibbosa Boonea seminuda Odostomia laevigata Odostomia producta Eulimastoma teres Sayella fusca Sayella hemphilli Turbonilla interrupta Turbonilla conradi Turbonilla cf. dalli Turbonilla hemphilli Eulimella smithii Lephalapsidea sp.

Family Acteonidae

Rictaxis punctostriatus

Family Cylichnidae

Acteocina canaliculata Tornatina inconspicua

Family Haminoeidae

Haminoea succinea Haminoea antillarum

Class Bivalvia

Family Nuculidae *Nucula crenulata*

FamilyArcidae *Anadara transversa*

Family Mytilidae

Brachidontes exustus Amygdalum papyrium

Family Lucinidae

Parvilucina multilineata

Family Ostreidae *Crassostrea virginica*

Family Ungulinidae *Diplodonta semiaspera*

Family Lasaeidae

Orobitella floridana Mysella planulata Erycina floridana

Family Cardiidae

Laevicardium mortoni

Family Mactridae

Mulinia lateralis

Family Semelidae

Ervilia concentrica

Family Tellinidae

Macoma tenta Macoma constricta Tellina iris Tellina lineata Tellina versicolor Tellina alternata Tellina squamifera Tellina tampaensis

Family Solecurtidae

Tagelus plebeius Tagelus divisus

Family Semelidae

Abra aequalis Semele proficua

Family Veneridae

Dosinia discus Mercenaria campechiensis Pitar sp. Macrocallista nimbosa Anomalocardia auberiana Parastarte triquetra

Family Myidae *Sphenia antillensis*

FamilyCorbulidae

Corbula contracta

Family Lyonsiidae

Lyonsia floridana

Family Thraciidae *Asthenothaerus hemphilli*

Phylum Arthropoda **Class Malacostraca Order Leptostraca Family Nebaliidae** *Nebalia*

Order Mysidacea

Family Mysidae *Mysidopsis spp. Bowmaniella sp. Brasilomysis sp. Americamysis bahia*

Order Cumacea Family Leuconidae *Leucon americanus*

FamilyDiastylidae

Oxyurostylis smithi Oxyurostylis lecroyae

Family Nannastacidae *Almyracuma proximoculi*

Family Bodotriidae *Cyclaspis cf. varians*

Order Tanaidacea Family Kalliapseudidae *Kalliapseudes sp. A*

Family Leptocheliidae

Leptochelia sp.

Order Isopoda Family Anthuridae *Cyathura polita Ptilanthura tenuis*

Family Hyssuridae

Xenanthura brevitelson Amakusanthura magnifica

Family Sphaeromatidae *Eurydice personata*

Family Idoteidae

Erichsonella attenuata Edotia triloba

Order Amphipoda

Family Ampeliscidae *Ampelisca abdita Ampelisca vadorum Ampelisca agassizi Ampelisca holmesi Ampelisca sp. C Ampelisca sp. B Ampelisca sp. A*

Family Ampithoidae

Cymadusa compta

Family Aoridae

Paramicrodeutopus cf. myersi Rudilemboides naglei Bemlos sp.

Family Bateidae *Batea catharinensis*

Family Ischyroceridae

Cerapus sp. C Erichthonius brasiliensis

Family Aoridae *Grandidierella bonnieroides*

Family Gammaridae

Elasmopus laevis Gammarus mucronatus Melita elongata

Family Haustoriidae *Acanthohaustorius uncinus*

Family Corophiidae *Microprotopus shoemakeri*

Family Liljeborgiidae *Listriella barnardi*

Family Lysianassidae *Shoemakerella lowreyi*

Family Oedicerotidae *Hartmanodes nyei*

Family Phoxocephalidae *Metharpinia floridana Eobrolgus spinosus*

Family Platyischnopidae *Eudevenopus honduranus*

Family Pariambidae

Paracaprella tenuis Paracaprella pusilla

Order Decapoda Family Penaeidae *Rimapenaeus constrictus*

Family Palaemonidae *Periclimenes americanus*

Family Alpheidae *Alpheus sp. Automate sp.*

Family Hippolytidae *Hippolyte sp.*

Family Processidae

Processa hemphilli Ambidexter symmetricus

Family Paguridae

Pagurus longicarpus Pagurus gymnodactylus Pagurus maclaughlinae

Family Porcellanidae

Polyonyx gibbesi

Family Upogebiidae *Upogebia affinis*

Family Leucosiidae

Persephona mediterranea

Family Panopeidae

Hexapanopeus angustifrons Panopeus bermudensis Rhithropanopeus harrisii Dyspanopeus texanus

Family Pinnotheridae

Pinnixa chaetopterana Pinnixa cf. pearsei Pinnixa A Pinnixa D

Phylum Sipuncula

Sipuncula sp. A

Phylum Phoronida *Phoronis ?architecta*

Phylum Brachipoda

Glottidia pyramidata

Phylum Echinodermata **Class Ophiuroidea Family Ophiactidae** *Hemipholis elongata*

Family Amphiuridae

Amphiodia nr. riisei Amphipholis squamata Amphipholis gracillima Ophiophragmus filograneus Amphioplus abditus Amphioplus thrombodes Amphioplus A Micropholis sp. Amphipholis atra

Class Echinoidea

Mellita tenuis

Class Holothuroidea

Holothuroidea C

Family Synaptidae

Synaptidae A Synaptidae C Leptosynapta

Phylum Hemichordata **Class Enteropneusta**

Enteropneusta B

Family Harrimaniidae

Stereobalanus canadensis

Class Cephalochordata Order Amphioxi Family Branchiostomidae *Branchiostoma floridae*

APPENDIX D SIMPER ANALYSES: COMPARISONS OF HILLSBOROUGH BAY CLUSTERS, 1993-1998 (TAXA EXPLAINING 25% OF DISSIMILARITY)

Groups B1 & B2a1

Average dissimilarity = 25.23

Groups B1 & B2b1

Average dissimilarity = 23.97

Groups B2a1 & B2b1

Average dissimilarity = 21.70

Groups B1 & B2a2

Average dissimilarity = 27.58

Groups B2a1 & B2a2

Average dissimilarity = 26.34

Groups B2b1 & B2a2

Average dissimilarity = 19.50

Groups B1 & B2b2

Average dissimilarity = 24.91

Groups B2a1 & B2b2

Average dissimilarity = 22.51

Groups B2b1 & B2b2

Average dissimilarity = 16.65

Groups B2a2 & B2b2

Average dissimilarity = 20.60

Groups B1 & A2

Average dissimilarity = 23.34

Groups B2a1 & A2

Average dissimilarity = 21.25

Groups B2b1 & A2

Average dissimilarity = 10.69

Groups B2a2 & A2

Average dissimilarity = 17.95

Groups B2b2 & A2

Average dissimilarity = 14.28

Groups B1 & B2

Average dissimilarity = 23.55

Groups B2a1 & B2

Average dissimilarity = 22.24

Groups B2b1 & B2

Average dissimilarity = 22.55

Groups B2a2 & B2

Average dissimilarity = 27.97

Groups B2b2 & B2

Average dissimilarity = 23.85

Groups A2 & B2

Average dissimilarity = 22.05

Groups B1 & A1

Average dissimilarity = 23.06

Groups B2a1 & A1

Average dissimilarity = 20.73

Groups B2b1 & A1

Average dissimilarity = 9.58

Groups B2a2 & A1

Average dissimilarity = 17.66

Groups B2b2 & A1

Average dissimilarity = 13.54

Groups A2 & A1

Average dissimilarity = 3.03

Groups B2 & A1

Average dissimilarity = 21.56

APPENDIX E SIMPER ANALYSES: COMPARISONS OF HILLSBOROUGH BAY BENTHIC ASSEMBLAGES, BY YEAR: 1993-1998 (TAXA EXPLAINING 10% OF DISSIMILARITY)

Groups 94 & 93 Average dissimilarity = 18.74

Groups 95 & 94 Average dissimilarity = 13.81

Groups 96 & 95 Average dissimilarity = 16.06

Appendix E- (continued)

Groups 96 & 97 Average dissimilarity = 20.04

Groups 97 & 98 Average dissimilarity = 17.83

